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可用于表征可电离化合物离子化影响的描述符研究进展

杨先海,刘会会*,王连军

南京理工大学环境与生物工程学院,江苏省化工污染控制与资源化高校重点实验室,南京 210094 收稿日期:2020-11-04 录用日期:2020-12-16

摘要:在人为有意生产的化学品或无意识产生的化学品中,可电离有机化合物(IOCs)均占有较大比重。在环境水体、生理或 实验 pH 条件下,IOCs 可解离为分子和离子形态。研究表明,IOCs 的分子和离子形态均对其表观物理化学、环境归趋和行为、 生态和健康毒性效应参数具有不可忽视的影响,因而在开展 IOCs 相关实验或理论研究时不应忽略离子化的影响。在构建 IOCs 相关预测模型时,核心是如何表征离子化的影响。本文从描述符入手,总结了可用于表征 IOCs 离子化影响的 4 类描述 符,即酸碱解离常数(pK_a)及其衍生参数(分子态和离子态的比例分数($\delta_{\partial f}$ 和 $\delta_{\delta f}$))、考虑离子化影响的分配系数包括正辛醇-水分布系数($logD_{ow}(pH)$)和进行形态修正的脂质体-水分配系数($logD_{lipw}(pH)$)、考虑离子参数的多参数线性自由能关系(离子描 述符 J^+ 和 J^-)、基于形态修正的量化参数,并展望了表征 IOCs 离子化影响的未来研究重点。

关键词:可电离有机化合物;离子化;离子形态;分子形态;分配参数;水生生物毒性;描述符;量化描述符;定量结构-活性/属性 关系

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Progress in Descriptors Used to Correct Influence of Ionization for Ionizable Organic Chemicals

Yang Xianhai, Liu Huihui^{*}, Wang Lianjun

Jiangsu Key Laboratory of Chemical Pollution Control and Resources Reuse, School of Environmental and Biological Engineering, Nanjing University of Science and Technology, Nanjing 210094, China Received 4 November 2020 accepted 16 December 2020

Abstract: Ionogenic organic chemicals (IOCs) are organic compounds with one or more ionizable function groups in their molecular structures. A large fraction of artificial chemicals or unintentional production chemicals are IOCs. Under the environmental, physiological and experimental pH condition, the IOCs may dissociate and exist as a mixture of neutral and ionized forms. It had been well documented that the neutral and ionized species of IOCs indeed had distinct physicochemical properties, environmental fate and behavior, ecological and health toxic effects. The observed parameters, properties, or endpoints influenced by ionization include but not limited to partition coefficients, photolysis rate constant, rate constants of hydroxyl radical, the adsorption capability to zeolite, bioconcentration, aquatic toxicity on fish, *Daphnia magna*, algae, *Tetrahymena pyriformis*, protein binding interaction, and so

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第一作者:杨先海(1985—),男,博士,研究方向为计算毒理学,E-mail: xhyang@njust.edu.cn

^{*} 通讯作者(Corresponding author), E-mail: hhliu@njust.edu.cn

on. In addition, the previous studies also implied that both forms of IOCs may contribute to their observed aforementioned apparent parameters. Thus, ionization should not be ignored in the related experimental and theoretical research of IOCs. Hitherto, how to correct the influence of ionization is one of the critical issues in deriving the predictive models (e.g. (quantitative) structure-activity/property relationship ((Q)SA/PR)) for IOCs. In this work, the available descriptors could be used to describe the effect of ionization for IOCs in the modeling were reviewed and summarized. Those descriptors include acid dissociation constant (pK_a) and derived parameters (the fractions of the neutral (δ_M) and ionized species (δ_1) at a given pH), distribution coefficient (the *n*-octanol/water distribution coefficient ($\log D_{ow}(pH)$) and speciation-corrected liposome-water distribution ratios ($\log D_{lip/w}(pH)$)), ionic descriptors in polyparameter linear free energy relationship (pp-LFER) equation (ionic descriptors J^+ and J^-), chemical form adjusted quantum chemical descriptors. Further investigations in correcting the ionization of IOCs were further discussed. **Keywords**: ionogenic organic chemicals (IOCs); ionization; ionized form; neutral form; partition coefficients; aquatic toxicity; descriptors; quantum chemical descriptor; (quantitative) structure-activity/property relationship

可电离有机化合物(ionizable organic chemicals, OH)、羧基(--COOH)、磺酸基(--SO,H)和氨基(--NH,)等)的有机酸碱、有机两性离子和有机盐等。在 人为有意生产的化学品或无意识产生的化学品中, IOCs均占有较大比例。据估计,在欧盟《化学品的 注册、评估、授权和限制法规》(REACH 法规)中登记 注册的14多万种化学品中,约半数属于 IOCs^[1-3]; 在目前使用的药物中,约80%属于 IOCs^[4],而85% ~95%的原料药也属于 IOCs^[5]。在水消毒副产物等 无意识产生的化学品中,也存在大量的取代有机羧 酸类、取代酚类等 IOCs^[6]。由于 IOCs 在生产、生活 中的大量使用,以及在水消毒等过程中无意识的产 生,导致其会通过多种途径进入各环境介质。人群、 环境生物也会通过经皮、经口和呼吸等多种途径暴 露于 IOCs,进而引发潜在的健康和/或生态危害效 应。因此,有必要从人工化学品或无意识产生的化 学品中筛选识别具有潜在健康和/或生态危害效应 的 IOCs,并对其采取适当的管控措施,以期保护人 群健康和生态安全[7]。

与不可电离化合物相比,IOCs 的重要特点是在环境水体、生理或实验不同 pH 条件下,会解离,从 而以不同比例的分子和离子形态共存。分子和离子 形态存在比例取决于其本身的酸碱解离常数(pK_a) 和环境/生理/试验 pH 条件(图 1)。式(1)显示了一元 酸碱解离程度的计算方程[®]:

$$\delta_{\dot{\gamma}\vec{\gamma}} = \frac{[M]}{[M]+[I]} = \frac{1}{1+10^{(pH-pK_{a})\cdot I_{ab}}}$$
$$\delta_{\vec{\beta}\vec{\gamma}} = \frac{[I]}{[M]+[I]} = \frac{10^{(pH-pK_{a})\cdot I_{ab}}}{1+10^{(pH-pK_{a})\cdot I_{ab}}}$$
(1)

 $δ_{\beta + 7}$ 和 $δ_{\beta + 7}$ 分别是分子态和离子态的比例分数; I_{ab} 是酸和碱指示系数, 酸和碱的 I_{ab} 分别取1和-1。 对 IOCs 而言,存在如下2个问题需要回答:一是其 共存的分子和离子形态是否具有不同的物理化学、 环境归趋和行为、生态和健康毒性效应?二是哪个 形态对化合物的表观属性和效应贡献更大?



注:HA 和 A⁻分别代表分子态和离子态;化合物 pK_a 实验数据 来源于 EPI Suit 4.1[™] 的 PhysProp 数据库^[9]。 Fig. 1 Species distribution of pentachlorophenol, 2,3-dichlorophenol and 3-chlorophenol



过去数十年的研究表明,IOCs的不同形态确实 表现不同的物理化学、环境归趋和行为、生态和健康 毒性效应^[10-14]。受到离子化影响的参数包括但不仅 限于分配参数(包括有机碳-水分配系数^[15-16]、牛血 清白蛋白-水分配系数[17-18]、肌肉蛋白-水分配系 数[19-20]、脂质体-水分配系数[3,21-23])、光解速率常 数[24-25]、羟基自由基速率常数[26]、沸石的吸附能 力[27]、生物富集[28]、鱼、大型溞、绿藻和梨形四膜虫等 水生生物的急性毒性[29-32]、蛋白相互作用[33-34]等。 相关研究已证实,部分 IOCs 的表观属性或毒性效 应值,受其分子态对应的属性或毒性效应值控制,如 生物富集、水生生物毒性效应等参数[32,35-37]:而有的 则取决于其离子态对应的属性或毒性效应值。我们 前期的研究表明,在 IOCs 与人运甲状腺素蛋白(hT-TR)相互作用的过程中,阴离子形态的酚类化合物、 全氟/多氟类化合物与 hTTR 的亲和力强于对应的 分子态^[34,38]。Goss 等^[39]的研究也表明,有机酸的生 物分配系数主要受阴离子控制。因此,现有的研究 结果,一方面表明 IOCs 的表观物理化学、环境归趋 和行为、生态和健康毒性效应参数确实受其离子化 的影响,这意味着在开展 IOCs 相关的实验或理论 研究时,不应忽视离子化的影响^[40];二是部分 IOCs 的表观属性或毒性效应参数值虽然取决于分子或离 子态的相应参数值,但不可否认的是仅考虑分子态 或离子态贡献时,往往会导致低估其属性或毒性效 应参数值^[41],这说明在进行 IOCs 相关的研究中,应 同时考虑其分子态和离子态的贡献^[42]。

目前,无论是人为有意生产的化学品,亦或无意 识产生的化学品,大部分仍缺乏基本的物理化学、环 境归趋和行为、生态和健康毒性效应参数数据^[43-45]。 而由于实验方法面临成本高、耗时长等问题,很难对 所有化学品一一进行实验测试。为了应对和解决该 问题,欧美等国家及经济合作与发展组织(OECD)等 国际组织都积极倡导使用预测技术来进行化学品优 先级设定、填补数据缺失等[46-50],如(定量)结构活性/ 属性关系((Q)SA/PR)模型。在过去几十年里,科学 家构建了许多能够预测化学品物理化学、环境归趋 和行为、生态和健康毒性效应参数的定量、定性预测 模型/软件/工具或专家系统。但是研究发现,绝大部 分预测模型/软件/工具或专家系统并不适用于 IOCs^[43,51-53]。一方面的原因是这些预测工具或专家 系统的模型应用域不含有 IOCs^[54-55];二是现有预测 模型选用的描述符未考虑离子化的影响,而仅是基 于 IOCs 分子态的描述符而构建的。而且仅采用 IOCs 分子态的描述符,有时很难构建可接受的模 型。例如, Endo 课题组在研究 IOCs 与牛血清蛋白 之间的分配系数(K_{BSAW})时,就发现常规描述符不 能构建具有较好预测能力的 IOCs K_{BSAW} 模型^[18]。因此,为了构建能够涵盖 IOCs 的预测模型,需要从两方面着手,一是在选择建模化学品时,需涵盖 IOCs;二是选用能够表征离子化影响的描述符进行建模。

考虑到描述符是预测模型构建的核心要素之 一^[43,50],选用合适的描述符来表征离子化的影响,对 于构建能预测 IOCs 物理化学、环境归趋和行为、生 态和健康毒性效应参数的模型具有重要意义。基于 此,本文总结了可用于表征可电离化合物离子化影 响的描述符,即酸碱解离常数及其衍生参数、考虑离 子化影响的分配系数、考虑离子参数的多参数线性 自由能关系、基于形态修正的量化参数,对其主要特 点进行了分析,并提出了研究展望。

1 酸碱解离常数及其衍生参数(Acid dissociation constant and derived parameters)

根据 Henderson-Hasselbalch 方程(式 2), 给定化 合物的 pK_a 值取决于其在平衡条件下分子态和离 子态的浓度^[56-57]。在某一特定 pH 条件下,具有较 小 pK_a 值的化合物,在平衡溶液中存在更大比例的 离子态浓度(式 3)。

$$pK_{a} = -\log K_{a} = -\log \frac{[A^{-}][H^{+}]}{[HA]} = -\log \frac{[A^{-}]}{[HA]} - \log[H^{+}] = -\log \frac{[A^{-}]}{[HA]} + pH$$
(2)

$$\frac{[A^-]}{[HA]} = 10^{pH-pK_a}$$
(3)

式中:[HA]、[A⁻]和[H⁺]分别是酸性物质平衡条件下 对应的分子态、阴离子和氢离子浓度。

从物理意义来看, pK_a 值可用于表征给定化合物在特定 pH 条件下的离子化状态^[58]。因此, 在建模中可采用 pK_a 来表征化合物离子化影响^[59-60]。自20世纪 60 年代 Fujita^[61]开始使用 pK_a 以来, pK_a 就常被作为预测变量来表征离子化的影响^[29,62-67]。例如, Schultz 等^[67]在构建酚类化合物对梨形四膜虫急性毒性预测模型时, 当采用 pK_a 作为第 2 个预测变量时(式(4)和(5)), 模型的决定系数(R^2_{ijijkk})从 0.783提高到 0.843。

$$-\log IGC_{50} = -0.772 + 0.627 \log K_{\rm OW}$$
(4)

 $-\log IGC_{50} = -0.120 + 0.614 \log K_{\rm OW} - 0.077 p K_{\rm a} \quad (5)$

 $n_{\text{illifiely}} = 54, R_{\text{illifiely}}^2 = 0.843, s = 0.315$

 $-\log IGC_{50}$ 是梨形四膜虫急性毒性效应值; $\log K_{ow}$

是正辛醇-水分配系数; n_{训练集}是训练集化合物数量; *s* 是标准误差。

使用该方法表征离子化影响时,前提是能获取 准确的 pKa值。然而,部分化合物却很难准确测定 其 pK_a 值。例如, 全氟辛酸的实验 pK_a 值介于 1.0 ~3.8^[68],其他全氟/多氟化合物的 pK。值也存在类 似问题^[69]。对于无法获取准确 pKa 值的 IOCs,该如 何处理呢?在这种情况下,可采用分子态(δ_{α_z})和离 子态(δ_{a+})的比例分数作为预测变量,而替代 pK 值。根据式(1),化合物 $\delta_{\beta\gamma}$ 和 $\delta_{\beta\gamma}$ 的值取决于环境/ 生理/实验 pH 值和其 pK_a。在一定程度上,使用 $\delta_{\alpha\gamma}$ 和 $\delta_{\alpha\gamma}$ 可减少因 pKa 变化带来的偏差。比如, 在生理 pH=7.40 的条件下, pKa 值介于 1.0~3.8 的 全氟辛酸δ_{高子}值取值均为1。目前,已有很多模型 采用 $\delta_{\alpha\gamma}$ 和 $\delta_{\alpha\gamma}$ 作为预测变量。例如,梨形四膜虫、 发光菌、大型溞和鲤鱼急性毒性[32,70-72]、土壤-水分配 系数[73]、溶解有机质-水分配系数[74]等参数预测模 型。例如,Qin 等^[70]采用 $\delta_{\beta\beta}$ 作用预测变量构建了 能预测中性分子、IOCs 对发光菌(V. fischeri) (式 6)、 大型溞(D. magna)(式 7)和鲤鱼(式 8)急性毒性的预 测模型:

 $-\log EC_{50}$ (发光菌)=1.04+0.701 $\log K_{OW}$ +1.11S+ 1.12 I_{NO_2} -0.157 $\log \delta_{3ff}$ (6)

 $n_{iiiiskg} = 102, R^2_{iiiiskg} = 0.790, s = 0.400$ $-\log EC_{50}$ (大型溞) = 1.73 + 0.628 log K_{ow} + 0.772 S + 0.899 I_{NO_2} + 0.542 log $\delta_{\beta\gamma\gamma}$ (7)

 $n_{ijjkk} = 102, R_{ijjkk}^2 = 0.790, s = 0.400$

 $-\log EC_{50}(\underline{\text{#}}\underline{\text{}}\underline{\text{}}\underline{\text{}}) = 1.52 + 0.581 \log K_{\text{OW}} + 0.767 S + 1.33 I_{\text{NO}_2} + 0.461 \log \delta_{\underline{\%}\underline{7}}$ (8)

$$n_{iii:44} = 102, R^2_{iii:44} = 0.790, s = 0.400$$

 $-\log EC_{50}$ (发光菌)、 $-\log EC_{50}$ (大型溞)和 $-\log EC_{50}$ (鲤 鱼)分别是发光菌、大型溞和鲤鱼急性毒性效应值;*S* 是化合物极化性参数;*I*_{NO2} 是分子中硝基个数。在 构建全氟和多氟化合物的 hTTR 干扰效应预测模型 时,我们选取了 δ_{sr} 值作为预测变量(式 9)^[38]:

 $logRP = -5.91 + 2.27HATS6m + 0.898\delta_{\mbox{\scriptsize B}\mbox{\scriptsize F}} - 1.94 \, qO_{adj}^{-}$ (9)

 $n_{ill \mbox{\tiny 1}\mbox{\tiny 1}\mbox{\tiny 1}\mbox{\tiny 1}\mbox{\tiny 2}\mbox{\tiny 1}\mbox{\tiny 2}\mbox{\tiny $$

 $n_{\text{验证集}} = 9, Q^2_{\text{验证集}} = 0.654, \text{RMSE}_{\text{验证集}} = 0.874$ logRP 是化合物与 hTTR 的相互作用势(logRP); HATS6m 是分子质量加权的杠杆自相关指数; qO^-_{adi} 是形态修正的分子中最负氧原子电荷; n_{验证集}是验证 集化合物数量; Q²_{CUM} 是模型所提取的所有 PLS 主 成分所能解释的因变量总方差的比例; Q²_{验证集}是验 证集外部可解释方差; RMSE_{训练集}和 RMSE_{验证集}代表 训练集和验证集均方根误差; P 为显著性水平。

 pK_a 实验值可从文献或数据库查询,如 eChemPortal (https://www.echemportal.org/echemportal/), Drugbank (https://www.drugbank.ca/), ChemIDplus (https://chem.nlm.nih.gov/chemidplus/), Physprop (http://esc. syrres. com/fatepointer/search. asp), OECD QSAR toolbox (https://qsartoolbox.org/)等。若无实验 pK_a 值,可采用 ChemAxon (http://www.chemaxon. com), Virtual Computational Chemistry Laboratory (http://www.vcclab.org/lab), SPARC (http://www.archemcalc.com/sparc.html), SciFinder (https://scifinder.cas. org)等软件预测。

2 考虑离子态贡献的分配系数(Distribution coefficient with ionization correction)

logKow 可表征不可电离化合物或 IOCs 的分子 态从水相分配到有机相的能力,因而从20世纪60 年代开始^[75], log Kow 就被用来预测各种涉及分配的 环境归趋和行为、生态和健康毒性效应参数^[76]。例 如, EPI Suit 4.1[™]的 ECOSAR[™] 模块主要通过 logKow 来预测化合物鱼、大型溞和绿藻急慢性毒性 效应数据。但是,在涉及 IOCs 的场合, $\log K_{ow}$ 的预 测能力可能变差。例如,图2(a)显示了在 pH=6.0、 7.8 和 9.0 条件下, log Kow 与 IOCs 对大型 潘 24 h 急 性毒性数据(-logEC₅₀)之间的关系。由该图可知, logK_{ow}与-logEC₅₀之间不存在显著线性相关性(pH =7.8 和 9.0)或仅存在较弱的相关性(pH=6.0)。如何 提高二者之间的相关性呢? 一般认为,考虑离子态 贡献的正辛醇-水分布系数($\log D_{ow}(pH)$)比 $\log K_{ow}$ 更适合 IOCs^[77-78]。如图 2(b)所示,在 pH=6.0、7.8 和 9.0 条件下, log Dow(pH)与-log EC50 的 Pearson 线性 相关系数分别从与 logKow 的 0.350 (P<0.01)、0.250 (P>0.05)和 0.174 (P>0.05)提高到 0.678 (P<0.0001)、 0.798 (P<0.0001)和 0.845 (P<0.0001)。这进一步说 明了考虑离子化的重要性。

 $log D_{ow}(pH)$ 可以通过式(10)或(11)计算: $log D_{ow}(pH) = log K_{ow} - log(1+10^{pH-pK_a})$ (10)

$$\log D_{\rm ow}(\rm pH) = \frac{\sum_{i=1}^{n} [\rm species]_{i,\rm octanol}}{\sum_{i=1}^{n} [\rm species]_{i,\rm water}}$$
(11)

式中:*i*指分子或离子态; [species]_{*i*,octanol}和 [species]_{*i*,water}分别指第*i*种形态在正辛醇和水中的浓度。目前,log D_{ow} (pH)已被广泛用于预测可电离环境污染物的相关属性或毒性效应^[32,79-82]或可电离药物分子相关参数^[12,78,83]。例如,Ou等^[80]基于log D_{ow} (pH)构建了预测 IOCs 鱼类肌肉蛋白-水分配系数(log $K_{MP/w}$)的预测模型(式 12):



$$\log K_{\rm MP/w} = -0.715 + 0.743 \log D_{\rm ow} \,(\rm pH = 7.0) + 0.0604 \, n_{\rm Car} \tag{12}$$

 $n_{\text{验证集}} = 11, Q_{\text{验证集}}^2 = 0.915, \text{RMSE}_{\text{验证集}} = 0.244$ n_{Car} 是分子结构中 sp2 杂化的芳香碳原子个数; Q_{Loo}^2 是去一法交叉验证系数。



图 2 $\log K_{ow}$, $\log D_{ow}$ (pH)与 $-\log EC_{50}$ 的关系

注:logK_{ow} 表示正辛醇-水分配系数,logD_{ow}(pH)表示正辛醇-水分布系数;可电离有机化合物(IOCs)的大型溞急性毒性数据(-logEC₅₀, 24 h)和 相应的 logK_{ow} 数据来源于 Li 等^[84];logD_{ow}(pH)数据采用 MarvinSketch 15.6.29.0, 2015(ChemAxo, http://www.chemaxon.com)软件预测。 Fig. 2 Correlation between logK_{ow}, logD_{ow}(pH) and -logEC₅₀

Note: $\log K_{OW}$ is *n*-octanol/water partition coefficient, and $\log D_{OW}$ is *n*-octanol/water distribution coefficient; the acute toxicity data ($-\log EC_{50}$, 24 h) of ionizable organic chemicals (IOCs) to *Daphnia magna*, corresponding $\log K_{OW}$ values were obtained from Li et al^[84], and $\log D_{OW}(pH)$ values of those IOCs were predicted employing MarvinSketch 15.6.29.0, 2015 (ChemAxon, http://www.chemaxon.com).

从结构上看,测定 $\log K_{ow}$ 和 $\log D_{ow}(pH)$ 使用 的辛醇是均相体系,而真实的生物膜含有磷脂双分 子层,属于非均相体系。这意味着化合物在辛醇相 的分配行为与其在真实生物膜中的分配或跨膜行为 具有较大的差异。因此,相比于 $\log K_{ow}$ 或 $\log D_{ow}$ (pH),使用膜-水分配系数($\log K_{m/w}$)来表征化合物的 膜通透性或膜累积能力具有更大的优势。然而,由 于真实的生物膜很难获取及开展实际的测试,一般 采用脂质体-水分配系数($\log K_{lip/w}$)来近似替代 $\log K_{m/w}$ ^[85]。对于不可电离化合物,可以采用分子态 的 $\log K_{lip/w-\beta +}$ 作为预测变量。但是对于 IOCs,需要 使用在特定 pH 条件下进行形态修正的脂质体-水 分配系数(($\log D_{lip/w}(pH)$)来表征。其定义如下^[86]:

 $\log D_{\text{lip/w}}(\text{pH}) = \log K_{\text{lip/w-} \text{H} \neq -i} \cdot \delta_{\text{H} \neq -i} + \sum_{i=1}^{n} \log K_{\text{lip/w-} \text{B} \neq -i}$ $\cdot \delta_{\text{B} \neq -i}$ (13)

式中: $\log K_{lip/w-\beta +}$ 和 $\log K_{lip/w-\beta + i}$ 分别是 IOCs 分子 态和第 *i* 种离子态的脂质体-水分配系数; $\delta_{\beta + i}$ 是 第 *i* 种离子态的比例分数。目前,已有较多模型采 用 $\log D_{lip/w}(pH)$ 作为预测变量预测 IOCs 的相关属性 或毒性。如,细菌毒性^[87-88]、藻毒性^[22]、斑马鱼胚胎 毒性^[10,89]、发光菌生物发光抑制毒性^[90]、生物累积 性^[91-92]和吸附属性^[93]等。例如,Klüver 等^[89]采用 $\log D_{lip/w}(pH)$ 来预测 IOCs 的斑马鱼胚胎急性毒性参 数($-\log LC_{so}$)(式 14);

 $-\log LC_{50} = -2.22 + 0.99 \log D_{lip/w}$ (pH) (14) 化合物的 $\log K_{ow}$ 和 $\log D_{ow}$ 值可以从文献、数据 库查询得到,也可通过软件预测。如 EPI Suit 4.1TM、 VEGA (https://www.vegahub.eu/)、OECD QSAR toolbox (https://qsartoolbox.org/)、ChemAxon (http://www. chemaxon.com)等。 $\log K_{lip/w}$ 和 $\log D_{lip/w}$ (pH)数据可 通过查阅文献获取或根据文献报道的模型进行预 测。例如,在前期的研究中,我们查询了 290 种化合 物的分子态脂质体-水分配系数($\log K_{lip/w-分子}$)、106 种 化合物的离子态脂质体-水分配系数($\log K_{lip/w-\beta7}$)和 306 种化合物进行形态修正的脂质体-水分配系数 (($\log D_{lip/w}$ (pH))数据,同时构建了能预测 $\log K_{lip/w-分7}$ 、 $\log K_{\lim/w-g}$ 和 $\log D_{\lim/w}$ (pH)数据的模型^[85]。

3 考虑离子参数的多参数线性自由能关系(Ionic descriptors in polyparameter linear free energy relationship (pp-LFER) equation)

传统的多参数线性自由能关系(pp-LFER)模型 可用于预测中性化合物从水相/气相到各种有机相 的分配参数^[94]。pp-LFER 模型一般采用如下 3 个方 程来表征^[95-97]:

$$SP = c + eE + sS + aA + bB + vV \tag{15}$$

$$SP = c + eE + sS + aA + bB + lL \tag{16}$$

$$SP = c + sS + aA + bB + lL + vV \tag{17}$$

式中:SP一般指分配系数; *E* 是化合物过量摩尔折 射率; *S* 是化合物极化性参数; *A* 和 *B* 是分子整体氢 键酸度和碱度; *V* 是 McGowan 分子体积; *L* 是 298 K 条件下,正十六烷-空气分配系数的对数值; *c* 是常 数项; *e*、*s*、*a*、*b* 和 *v* 是系数。目前,该方程在预测不 可电离化合物的相关分配系数方面得到了较多的应 用,例如在 Web of Science 数据库,通过"polyparameter linear free energy relationship"作为关键词,可以 检索到数十篇相关论文,在这里我们就不详细列出 相关应用。为了适用于 IOCs, Abraham 和 Zhao^[98]在 传统 pp-LFER 方程基础上通过引入了 2 个新的离 子描述符,即 J^+ 和 J^- ,提出了考虑离子参数的 pp-LFER 方程:

SP=*c*+*eE*+*sS*+*aA*+*bB*+*vV*+*j*⁺*J*⁺+*j*⁻*J*⁻ (18) 通过采用考虑离子参数的 pp-LFER 方程,前人 构建了可预测 IOCs 的多种参数,包括有机溶剂-水 分配系数^[99-101]、脂质体-水分配系数^[3]、血清白蛋白-水分配系数^[18]、肌肉蛋白水分配系数^[20]、活性炭-水 分配系数^[102]和针铁矿-溶剂分配系数^[103]等。例如, Henneberger 等构建了能够预测 IOCs 鸡类肌肉蛋 白-水分配系数(log*K*_{MP/w})的预测模型^[20]:

$$\log K_{\rm MP/w} = -0.24 + 0.68 E - 0.76 S - 0.20 A - 2.29 B + 2.51 V - 0.68 L^{+} + 2.89 L^{-}$$
(10)

$$2.51 V - 0.68 J^+ + 2.89 J^- \tag{19}$$

 $n_{\text{illifs} \#} = 86, R^2_{\text{illifs} \#} = 0.89, \text{RMSE}_{\text{illifs} \#} = 0.29$

pp-LFER 建模所需参数 $E_s S_s A_s B_s V$ 和 L 可通 过查询 UFZ-LSER 数据库(https://www.ufz.de/ index. php? en = 31698&contentonly = 1&m = 0&lserd_data [mvc]=Public/start)或大连理工大学陈景文教授团队 开发的在线程序预测所需参数(http://www.pplfer.online/)。离子描述符 J^+ 和 J^- 可通过查阅文献获取。 4 基于形态修正的量化参数(Chemical form adjusted quantum chemical descriptors)

化合物的量子化学描述符一般具有明确的物理 化学意义,有利于进行模型机理解释^[59,104]。为了预 测 IOCs 的相关属性或毒性效应,可对量化描述符 进行形态修正^[34,105]:

$$X_{\&_{\mathbb{T}}} = X_{\mathcal{H}\mathcal{F}} \cdot \delta_{\mathcal{H}\mathcal{F}} + \sum_{i=1}^{n} X_{\boxtimes \mathcal{F} - i} \cdot \delta_{\boxtimes \mathcal{F} - i}$$
(20)

式中: $X_{\&st}$ 、 $X_{\Im T}$ 和 $X_{\& g T-i}$ 分别是 IOCs 形态修正、分 子态和第*i*种离子态的量化描述符。通过采用形态 修正的量化描述符,可显著提高模型的预测质量。 例如,在构建 IOCs 牛血清白蛋白-水分配系数^[106]和 大型溞急性毒性效应参数($-\log EC_{so}$)^[8]时,我们比较 了仅采用分子态描述符和基于形态修正的描述符构 建的模型质量,发现采用后者构建的模型其 $R^2_{ijj \& s \#}$ 和 $Q^2_{\& iii \& \$}$ 分别从 0.508 和 0.220 提高到 0.707 和 0.581(牛血清白蛋白-水分配系数)、从 0.705 和 0.651 到 0.875 和 0.851(大型溞急性毒性效应)。构建的大 型溞急性毒性效应预测模型如下(式(21)和(22))^[8]:

 $-\log EC_{50}(pH=7.8) = 15.7 - 9.59 qH_{-\%7}^{+} - 9.12\tau_{-\%7} + 474 V_{s-\%7} + 25.6 E_{HOMO-\%7}$ (21)

 $n_{\text{ill} \text{is} \# =} 48$, $R_{\text{ill} \text{is} \#}^2 = 0.705$, $Q_{\text{LOO}}^2 = 0.622$, RMSE_{illis \#}=0.569, *P*<0.0001

 $n_{\text{\pmi}\ensuremath{\mathbb{I}}\ensuremath{\mathbb{R}}\ensuremath$

 $n_{\text{验证集}} = 15, Q_{\text{验证集}}^2 = 0.851, \text{RMSE}_{\text{验证集}} = 0.336$ 式中: $qH_{- \Im 7}^+$ 是分子态中氢原子的最正净电荷; $\tau_{. \Im 7}^-$ 是分子态静电势的平衡参数; $V_{s- \Im 7}^-$ 是分子态分子表面上静电势平均值; $E_{\text{HOMO-} \Im 7}^-$ 是分子态分子最高占据轨道能; $qD_{-\&_{\text{ET}}}^-$ 是形态修正的分子中电子供体原子电荷;polar_ $\&_{\&_{\text{ET}}}^-$ 是形态修正的分子极化率; $\Pi_{-\&_{\text{ET}}}^-$ 是形态修正的分子表面静电势的分散度。

在表1中列出了18种形态修正的量化描述符。 基于这些参数,我们已成功构建了IOCs与人运甲 状腺素蛋白(hTTR)亲和力^[34,38,107-110]、牛血清白蛋白-水分配系数^[106]、脂质体-水分配系数^[85]和大型溞急 性毒性^[8]等参数预测模型。从定义可知,基于形态 修正的量化参数具有2个方面的特点:(1)可以直接 根据IOCs的分子态和离子态结构计算得到;(2)可 以同时考虑分子态和共存的多种离子态的贡献。

14010 1	rioposed enemies	a form adjusted quantum energies asserptors used in the (Q)SFFFFF inducing
序号	描述符	描述
No.	Descriptors	Description
1	$qO_{Mate}^{-} qO_{Adi}^{-}$	形态修正的分子中最负氧原子电荷
	I ISIL I AU	The chemical form adjusted most negative net atomic charge on an oxygen atom
2	aX_{ikr}^{-} aX_{irr}^{-}	形态修正的分子中最负卤素原子电荷
	97、修止 97、Adj	The chemical form adjusted most negative net atomic charge on a halogen atom
3	aD_{m-1} aD_{m-1}	形态修正的分子中电子供体原子电荷
	9D修止 9D Adj	The chemical form adjusted most negative net atomic charge on an electron donor atom
4	$V \rightarrow V$	形态修正的分子表面最正静电势
	▼s, max 修止 ▼s, max-Adj	The chemical form adjusted most positive values of the molecular surface potential
5	$V \dots V$	形态修正的分子表面最负静电势
	▼s, min 修正 ▼s, min-Adj	The chemical form adjusted most negative values of the molecular surface potential
6	\mathbf{V}^+ \mathbf{V}^+	形态修正的分子表面正静电势的平均值
	V _s 修正 V _{sAdj}	The chemical form adjusted averages of the positive potentials on the molecular surface
7	17 - 17 -	形态修正的分子表面负静电势的平均值
	V _s 修正 V _{sAdj}	The chemical form adjusted averages of the negative potentials on the molecular surface
8	VV	形态修正的分子表面静电势平均值
	Vs;aver修正 Vs;aver-Adj	The chemical form adjusted average potentials on the molecular surface
9	пп	形态修正的分子表面静电势的分散度
	III _{修正} III _{Adj}	The chemical form adjusted average deviation of surface potential
10		形态修正的静电势平衡参数
	$ au_{ m eta_{ m E}} au_{ m Adj}$	The chemical form adjusted balance parameter of the surface potential
11	1. 1 1. 1	形态修正的分子偶极矩
	dipole _{修正} dipole _{Adj}	The chemical form adjusted molecular dipolemoment and its descriptor
12	, ,	形态修正的分子极化率
	polar _{修正} polar _{Adj}	The chemical form adjusted molecular polarizability
13	F F	形态修正的分子最高占据轨道能
	E _{HOMO 修正} E _{HOMO-Adj}	The chemical form adjusted highest occupied molecular orbital energy and its descriptor
14		形态修正的分子最低未占据轨道能
	E _{LUMO 修正} E _{LUMO-Adj}	The chemical form adjusted lowest unoccupied molecular
		orbital energy and its chemical form adjusted descriptor
15		形态修正的电负性指数
	$\omega_{$ 修正 $\omega_{ m Adj}}$	The chemical form adjusted electrophilicity index
16		形态修正的化学势
	$\mu_{ m \&E}\;\mu_{ m Adj}$	The chemical form adjusted chemical potential
17		形态修正的化学硬度
	$oldsymbol{\eta}_{ extsf{\&E}} oldsymbol{\eta}_{ extsf{Adj}}$	The chemical form adjusted chemical hardness
18	••	形态修正的分子体积
	$V_{$ 修正 $V_{ m Adj}$	The chemical form adjusted molecular volume

表1 基于形态修正的(Q)SA/PR 模型量化参数

Table 1 Proposed chemical form adjusted quantum chemical descriptors used in the (Q)SA/PR modeling

注:(Q)SA/PR 表示(定量)结构活性/属性关系。

Note: (Q)SA/PR is (quantitative) structure-activity/property relationship.

5 总结与展望(Conclusion and prospect)

IOCs的分子态和离子态对其物理化学、环境归 趋和行为、生态和健康毒性效应参数均具有不同贡 献。在开展 IOCs 相关实验和理论研究时,不应忽 视 IOCs 离子化的影响。截止目前,研究人员针对 在建模中如何考虑 IOCs 离子化影响的问题,提出

了4种可用于表征 IOCs 离子化影响的描述符,即酸 碱解离常数及其衍生参数、考虑离子化影响的分配 系数、考虑离子参数的多参数线性自由能关系、基于 形态修正的量化参数,并成功将其应用于 IOCs 各 种分配系数、水生毒性和蛋白结合效应等参数的 (Q)SA/PR 模型构建。

针对现有研究进展,对今后 IOCs 建模研究提 出了以下建议。(1) 预测指标方面:需要进一步识别 还有哪些参数受离子化的影响;(2) 描述符方面:需 要研究如何更准确获取所需描述符。如前所述,在 建模中要考虑离子化的影响,核心是获取 pK_{ax} logD_{ow}(pH)、logD_{lin/w}(pH)、J⁺和 J⁻、量化参数等相关 描述符。但是,在现有 IOCs 中,仅部分具有可靠的 pK_a 、log $D_{ow}(pH)$ 和 log $D_{lin/w}(pH)$ 等参数的实验数据。 例如, 仅几百种 IOCs 拥有 log D_{lin/w}(pH)实验数据^[85], 几千种 IOCs 有 pK_a 实验数据^[58]。因此,需要进一 步测定 IOCs 的 pKa、logDow(pH)和 logDin/w(pH)等参 数的实验数据:同时也可以构建更多能够准确预测 这些参数的模型工具。对量化参数而言,已引入了 18 种参数用于建模。根据 Mamy 等[111]的研究结 果,目前文献中已报道的量化参数已达248种。需 要进一步研究是否还存在其他量化参数适合用于进 行形态修正并用于建模。(3) (Q)SA/PR 模型构建方 面:针对受离子化影响的参数,构建更多预测模型, 同时考虑将其集成到现有软件工具中或开发新的软 件工具。(4) 非(Q)SA/PR 方法的开发与应用: 需积极 探索其他适合用于表征离子化影响的非(Q)SA/PR 方法。IOCs 的一些属性或毒性效应参数涉及小分 子与生物大分子(如生物膜、蛋白等)的相互作用。 而分子对接、分子动力学和耦合量子力学/分子力学 (QM/MM)等分子模拟方法常用于研究小分子与生 物大分子的相互作用。是否可以采用分子模拟方法 来表征 IOCs 离子化的影响呢? 在这方面,已有部 分研究进行探讨,例如,Bittermann 等^[3,112]采用 COS-MOmic 方法成功预测了 IOCs 生物膜-水分配系数。 我们采用 QM/MM 方法研究了酚类化合物与 hTTR 的相互作用,发现实验测定的酚类化合物与 hTTR 亲和力数据(logRP)与基于酚类化合物分子态计算 的结合能(E_{结合能-分子态})之间无显著性线性相关性。 但是,当采用形态修正的结合能(E_{结合能-形态修正})后,其 与 logRP 则存在显著性线性相关性^[34]。因此,需要 进一步探索采用分子模拟等非(Q)SA/PR 方法来预 测 IOCs 相关属性或毒性参数的可行性。

通讯作者简介:刘会会(1985—),女,博士,副教授,主要研究 方向为新型被动采样技术的研发与应用、环境中微塑料的环 境行为与毒理学效应研究、有机污染物生态毒理效应的计算 模拟研究等。

参考文献(References):

- Trapp S, Franco A, MacKay D. Activity-based concept for transport and partitioning of ionizing organics [J]. Environmental Science & Technology, 2010, 44(16): 6123-6129
- [2] Armitage J M, Arnot J A, Wania F, et al. Development and evaluation of a mechanistic bioconcentration model for ionogenic organic chemicals in fish [J]. Environmental Toxicology and Chemistry, 2013, 32(1): 115-128
- [3] Bittermann K, Spycher S, Goss K U. Comparison of different models predicting the phospholipid-membrane water partition coefficients of charged compounds [J]. Chemosphere, 2016, 144: 382-391
- [4] Manallack D T, Prankerd R J, Nassta G C, et al. A chemogenomic analysis of ionization constants-implications for drug discovery [J]. ChemMedChem, 2013, 8(2): 242-255
- [5] Karlsson M V, Carter L J, Agatz A, et al. Novel approach for characterizing pH-dependent uptake of ionizable chemicals in aquatic organisms [J]. Environmental Science & Technology, 2017, 51(12): 6965-6971
- [6] Liu X Y, Chen L, Yang M T, et al. The occurrence, characteristics, transformation and control of aromatic disinfection by-products: A review [J]. Water Research, 2020, 184: 116076
- [7] European Centre for Ecotoxicology and Toxicology of Chemicals (ECETOC). Environmental exposure assessment of ionizable organic compounds. Technical Report No. 123 [R]. Brussels: ECETOC, 2013
- [8] 席越,杨先海,张红雨,等.基于形态修正的描述符构 建可电离化合物对大型溞急性毒性的 QSAR 模型[J]. 生态毒理学报, 2019, 14(4): 183-191
 Xi Y, Yang X H, Zhang H Y, et al. Development of acute toxicity of *Daphnia magna* QSAR models for ionogenic organic chemicals based on chemical form adjusted descriptors [J]. Asian Journal of Ecotoxicology, 2019, 14(4): 183-191 (in Chinese)
- [9] United States Environmental Protection Agency (US EPA). Estimation Programs Interface Suite[™] for Microsoft[®] Windows, V 4.10 [CP]. Washington DC: US EPA, 2012
- [10] Bittner L, Klüver N, Henneberger L, et al. Combined iontrapping and mass balance models to describe the pH-de-

pendent uptake and toxicity of acidic and basic pharmaceuticals in zebrafish embryos (*Danio rerio*) [J]. Environmental Science & Technology, 2019, 53(13): 7877-7886

- [11] Henneberger L, Goss K U. Environmental Sorption Behavior of Ionic and Ionizable Organic Chemicals [J]. Reviews of Environmental Contamination and Toxicology, 2019, 253: 1-21
- [12] Manallack D T, Prankerd R J, Yuriev E, et al. The significance of acid/base properties in drug discovery [J]. Chemical Society Reviews, 2013, 42(2): 485-496
- [13] Strope C L, Mansouri K, Clewell H J III, et al. Highthroughput in-silico prediction of ionization equilibria for pharmacokinetic modeling [J]. Science of the Total Environment, 2018, 615: 150-160
- [14] 徐世积,何影,李思齐,等.环境中可电离有机化合物 生物有效性研究进展[J]. 生态与农村环境学报, 2017, 33(5): 385-395
 Xu S J, He Y, Li S Q, et al. Review of researches on bioavailability of ionizable organic compounds in environment [J]. Journal of Ecology and Rural Environment, 2017, 33(5): 385-395 (in Chinese)
- [15] Franco A, Fu W J, Trapp S. Influence of soil pH on the sorption of ionizable chemicals: Modeling advances [J]. Environmental Toxicology and Chemistry, 2009, 28 (3): 458-464
- [16] Tülp H C, Fenner K, Schwarzenbach R P, et al. pH-dependent sorption of acidic organic chemicals to soil organic matter [J]. Environmental Science & Technology, 2009, 43(24): 9189-9195
- [17] Endo S, Goss K U. Serum albumin binding of structurally diverse neutral organic compounds: Data and models [J]. Chemical Research in Toxicology, 2011, 24 (12): 2293-2301
- [18] Henneberger L, Goss K U, Endo S. Equilibrium sorption of structurally diverse organic ions to bovine serum albumin [J]. Environmental Science & Technology, 2016, 50 (10): 5119-5126
- [19] Endo S, Bauerfeind J, Goss K U. Partitioning of neutral organic compounds to structural proteins [J]. Environmental Science & Technology, 2012, 46(22): 12697-12703
- [20] Henneberger L, Goss K U, Endo S. Partitioning of organic ions to muscle protein: Experimental data, modeling, and implications for *in vivo* distribution of organic ions [J]. Environmental Science & Technology, 2016, 50(13): 7029-7036
- [21] Droge S T J, Hermens J L M, Gutsell S, et al. Predicting the phospholipophilicity of monoprotic positively charged amines [J]. Environmental Science Processes & Impacts,

2017, 19(3): 307-323

- [22] Neuwoehner J, Escher B I. The pH-dependent toxicity of basic pharmaceuticals in the green algae *Scenedesmus vacuolatus* can be explained with a toxicokinetic ion-trapping model [J]. Aquatic Toxicology, 2011, 101 (1): 266-275
- [23] Endo S, Escher B I, Goss K U. Capacities of membrane lipids to accumulate neutral organic chemicals [J]. Environmental Science & Technology, 2011, 45 (14): 5912-5921
- [24] Wei X X, Chen J W, Xie Q, et al. Distinct photolytic mechanisms and products for different dissociation species of ciprofloxacin [J]. Environmental Science & Technology, 2013, 47(9): 4284-4290
- [25] Xie Q, Chen J W, Zhao H X, et al. Different photolysis kinetics and photooxidation reactivities of neutral and anionic hydroxylated polybrominated diphenyl ethers [J]. Chemosphere, 2013, 90(2): 188-194
- [26] Luo X, Wei X X, Chen J W, et al. Rate constants of hydroxyl radicals reaction with different dissociation species of fluoroquinolones and sulfonamides: Combined experimental and QSAR studies [J]. Water Research, 2019, 166: 115083
- [27] Xie J, Meng W N, Wu D Y, et al. Removal of organic pollutants by surfactant modified zeolite: Comparison between ionizable phenolic compounds and non-ionizable organic compounds [J]. Journal of Hazardous Materials, 2012, 231-232: 57-63
- [28] Trapp S. Bioaccumulation of Polar and Ionizable Compounds in Plants [M]//Ecotoxicology Modeling. Boston, MA: Springer US, 2009: 299-353
- [29] Cronin M T D, Zhao Y H, Yu R L. pH-Dependence and QSAR analysis of the toxicity of phenols and anilines to *Daphnia magna* [J]. Environmental Toxicology, 2000, 15 (2): 140-148
- [30] He Q, Wang X H, Sun P, et al. Acute and chronic toxicity of tetrabromobisphenol A to three aquatic species under different pH conditions [J]. Aquatic Toxicology, 2015, 164: 145-154
- [31] Kamaya Y, Fukaya Y, Suzuki K. Acute toxicity of benzoic acids to the crustacean *Daphnia magna* [J]. Chemosphere, 2005, 59(2): 255-261
- [32] Zhao Y H, Yuan X, Su L M, et al. Classification of toxicity of phenols to *Tetrahymena pyriformis* and subsequent derivation of QSARs from hydrophobic, ionization and electronic parameters [J]. Chemosphere, 2009, 75(7): 866-871
- [33] Crisan M E, Bourosh P, Maffei M E, et al. Synthesis,

crystal structure and biological activity of 2-hydroxyethylammonium salt of *p*-aminobenzoic acid [J]. PLoS One, 2014, 9(7): e101892

- [34] Yang X H, Xie H B, Chen J W, et al. Anionic phenolic compounds bind stronger with transthyretin than their neutral forms: Nonnegligible mechanisms in virtual screening of endocrine disrupting chemicals [J]. Chemical Research in Toxicology, 2013, 26(9): 1340-1347
- [35] Nakamura Y, Yamamoto H, Sekizawa J, et al. The effects of pH on fluoxetine in Japanese medaka (*Oryzias latipes*): Acute toxicity in fish larvae and bioaccumulation in juvenile fish [J]. Chemosphere, 2008, 70(5): 865-873
- [36] Rendal C, Kusk K O, Trapp S. Optimal choice of pH for toxicity and bioaccumulation studies of ionizing organic chemicals [J]. Environmental Toxicology and Chemistry, 2011, 30(11): 2395-2406
- [37] Xing L Q, Liu H L, Giesy J P, et al. pH-dependent aquatic criteria for 2, 4-dichlorophenol, 2, 4, 6-trichlorophenol and pentachlorophenol [J]. Science of the Total Environment, 2012, 441: 125-131
- [38] Yang X H, Lyakurwa F, Xie H B, et al. Different binding mechanisms of neutral and anionic poly-/perfluorinated chemicals to human transthyretin revealed by in silico models [J]. Chemosphere, 2017, 182: 574-583
- [39] Goss K U, Bittermann K, Henneberger L, et al. Equilibrium biopartitioning of organic anions—A case study for humans and fish [J]. Chemosphere, 2018, 199: 174-181
- [40] Seward J R, Schultz T W. QSAR analyses of the toxicity of aliphatic carboxylic acids and salts to *Tetrahymena pyriformis* [J]. SAR and QSAR in Environmental Research, 1999, 10(6): 557-567
- [41] Boström M L, Berglund O. Influence of pH-dependent aquatic toxicity of ionizable pharmaceuticals on risk assessments over environmental pH ranges [J]. Water Research, 2015, 72: 154-161
- [42] Zhao Y H, Ji G D, Cronin M T D, et al. QSAR study of the toxicity of benzoic acids to *Vibrio fischeri*, *Daphnia magna* and carp [J]. Science of the Total Environment, 1998, 216(3): 205-215
- [43] Card M L, Gomez-Alvarez V, Lee W H, et al. History of EPI Suite[™] and future perspectives on chemical property estimation in US Toxic Substances Control Act new chemical risk assessments [J]. Environmental Science Processes & Impacts, 2017, 19(3): 203-212
- [44] Judson R, Richard A, Dix D J, et al. The toxicity data landscape for environmental chemicals [J]. Environmental Health Perspectives, 2009, 117(5): 685-695
- [45] 王中钰,陈景文,乔显亮,等.面向化学品风险评价的

计算(预测)毒理学[J]. 中国科学: 化学, 2016, 46(2): 222-240

Wang Z Y, Chen J W, Qiao X L, et al. Computational toxicology: Oriented for chemicals risk assessment [J]. Scientia Sinica (Chimica), 2016, 46(2): 222-240 (in Chinese)

- [46] Chen C C, Kuo D T F. Bioconcentration model for non-ionic, polar, and ionizable organic compounds in amphipod [J]. Environmental Toxicology and Chemistry, 2018, 37(5): 1378-1386
- [47] Worth A P, Bassan A, De Bruijn J, et al. The role of the European Chemicals Bureau in promoting the regulatory use of (Q)SAR methods [J]. SAR and QSAR in Environmental Research, 2007, 18(1-2): 111-125
- [48] Cronin M T D. (Q)SARs to predict environmental toxicities: Current status and future needs [J]. Environmental Science Processes & Impacts, 2017, 19(3): 213-220
- [49] Cronin M T D, Richarz A N, Schultz T W. Identification and description of the uncertainty, variability, bias and influence in quantitative structure-activity relationships (QSARs) for toxicity prediction [J]. Regulatory Toxicology and Pharmacology, 2019, 106: 90-104
- [50] Wang Z Y, Chen J W. Background, Tasks, Modeling Methods, and Challenges for Computational Toxicology [M]//Challenges and Advances in Computational Chemistry and Physics. Cham: Springer International Publishing, 2019: 15-36
- [51] European Commission. Technicalguidance document on risk assessment in support of Commission Directive 93/ 67/EEC on risk assessment for new notified substances, Commission Regulation (EC) No. 1488/94 on risk assessment for existing substances, Directive 98/8/EC of the European Parliament and of the Council concerning the placing of biocidal products on the market Part II. [S]. Brussel: European Communities, 2003
- [52] Trapp S, Schwartz S. Proposals to overcome limitations in the EU chemical risk assessment scheme [J]. Chemosphere, 2000, 41(7): 965-971
- [53] Huchthausen J, Mühlenbrink M, König M, et al. Experimental exposure assessment of ionizable organic chemicals in *in vitro* cell-based bioassays [J]. Chemical Research in Toxicology, 2020, 33(7): 1845-1854
- [54] Aalizadeh R, von der Ohe P C, Thomaidis N S. Prediction of acute toxicity of emerging contaminants on the water flea *Daphnia magna* by Ant Colony Optimization—Support Vector Machine QSTR models [J]. Environmental Science: Processes & Impacts, 2017, 19(3): 438-448
- [55] Vitale C M, Di Guardo A. A review of the predictive models estimating association of neutral and ionizable or-

ganic chemicals with dissolved organic carbon [J]. Science of the Total Environment, 2019, 666: 1022-1032

- [56] Katritzky A R, Kuanar M, Slavov S, et al. Quantitative correlation of physical and chemical properties with chemical structure: Utility for prediction [J]. Chemical Reviews, 2010, 110(10): 5714-5789
- [57] Todeschini R, Consonni V. Descriptors from Molecular Geometry [M]//Handbook of Chemoinformatics. Weinheim, Germany: Wiley-VCH Verlag GmbH, 2008: 1004-1033
- [58] Mansouri K, Cariello N F, Korotcov A, et al. Open-source QSAR models for pKa prediction using multiple machine learning approaches [J]. Journal of Cheminformatics, 2019, 11(1): 1-20
- [59] Zvinavashe E, Murk A J, Rietjens I M C M. Promises and pitfalls of quantitative structure-activity relationship approaches for predicting metabolism and toxicity [J]. Chemical Research in Toxicology, 2008, 21 (12): 2229-2236
- [60] Schaffer M, Licha T. A guideline for the identification of environmentally relevant, ionizable organic molecule species [J]. Chemosphere, 2014, 103: 12-25
- [61] Fujita T. The analysis of physiological activity of substituted phenols with substituent Constants1 [J]. Journal of Medicinal Chemistry, 1966, 9(6): 797-803
- [62] Lee Y G, Hwang S H, Kim S D. Predicting the toxicity of substituted phenols to aquatic species and its changes in the stream and effluent waters [J]. Archives of Environmental Contamination and Toxicology, 2006, 50(2): 213-219
- [63] Ramos-Nino M E, Clifford M N, Adams M R. Quantitative structure activity relationship for the effect of benzoic acids, cinnamic acids and benzaldehydes on *Listeria monocytogenes* [J]. The Journal of Applied Bacteriology, 1996, 80(3): 303-310
- [64] Nolte T M, Ragas A M J. A review of quantitative structure-property relationships for the fate of ionizable organic chemicals in water matrices and identification of knowledge gaps [J]. Environmental Science Processes & Impacts, 2017, 19(3): 221-246
- [65] Wayne Schultz T. The use of the ionization constant (pKa) in selecting models of toxicity in phenols [J]. Ecotoxicology and Environmental Safety, 1987, 14(2): 178-183
- [66] Schultz T W, Lin D T, Wesley S K. QSARs for monosubstituted phenols and the polar narcosis mechanism of toxicity [J]. Quality Assurance, 1992, 1(2): 132-143
- [67] Schultz T W, Bearden A P, Jaworska J S. A novel QSAR

approach for estimating toxicity of phenols [J]. SAR and QSAR in Environmental Research, 1996, 5(2): 99-112

- [68] Vierke L, Berger U, Cousins I T. Estimation of the acid dissociation constant of perfluoroalkyl carboxylic acids through an experimental investigation of their water-to-air transport [J]. Environmental Science & Technology, 2013, 47(19): 11032-11039
- [69] Goss K U. The pKa values of PFOA and other highly fluorinated carboxylic acids [J]. Environmental Science & Technology, 2008, 42(2): 456-458
- [70] Qin W C, Su L M, Zhang X J, et al. Toxicity of organic pollutants to seven aquatic organisms: Effect of polarity and ionization [J]. SAR and QSAR in Environmental Research, 2010, 21(5-6): 389-401
- [71] Zhao Y H, Zhang X J, Wen Y, et al. Toxicity of organic chemicals to *Tetrahymena pyriformis*: Effect of polarity and ionization on toxicity [J]. Chemosphere, 2010, 79(1): 72-77
- [72] Su L, Fu L, He J, et al. Comparison of *Tetrahymena pyriformis* toxicity based on hydrophobicity, polarity, ionization and reactivity of class-based compounds [J]. SAR and QSAR in Environmental Research, 2012, 23 (5-6): 537-552
- [73] Franco A, Trapp S. Estimation of the soil-water partition coefficient normalized to organic carbon for ionizable organic chemicals [J]. Environmental Toxicology and Chemistry, 2008, 27(10): 1995-2004
- [74] Vitale C M, Di Guardo A. Predicting dissolved organic carbon partition and distribution coefficients of neutral and ionizable organic chemicals [J]. Science of the Total Environment, 2019, 658: 1056-1063
- [75] Hansch C, Maloney P P, Fujita T, et al. Correlation of biological activity of phenoxyacetic acids with Hammett substituent constants and partition coefficients [J]. Nature, 1962, 194(4824): 178-180
- [76] 范莱文, 韦梅尔. 化学品风险评估[M]. 北京: 化学工业 出版社, 2010: 280-350
- [77] Scherrer R A, Howard S M. Use of distribution coefficients in quantitative structure-activity relations [J]. Journal of Medicinal Chemistry, 1977, 20(1): 53-58
- [78] Kah M, Brown C D. LogD: lipophilicity for ionisable compounds [J]. Chemosphere, 2008, 72(10): 1401-1408
- [79] Abbasitabar F, Zare-Shahabadi V. In silico prediction of toxicity of phenols to *Tetrahymena pyriformis* by using genetic algorithm and decision tree-based modeling approach [J]. Chemosphere, 2017, 172: 249-259
- [80] Ou W, Liu H H, He J Y, et al. Development of chicken and fish muscle protein—Water partition coefficients pre-

dictive models for ionogenic and neutral organic chemicals [J]. Ecotoxicology and Environmental Safety, 2018, 157: 128-133

- [81] Kah M, Brown C D. Prediction of the adsorption of ionizable pesticides in soils [J]. Journal of Agricultural and Food Chemistry, 2007, 55(6): 2312-2322
- [82] Wang Y Q, Liu H H, Yang X H, et al. Aquatic toxicity and aquatic ecological risk assessment of wastewater-derived halogenated phenolic disinfection byproducts [J]. Science of the Total Environment, 2021, Doi: 10.1016/j. scitotenv.2021.151089
- [83] Bayliss M K, Butler J, Feldman P L, et al. Quality guidelines for oral drug candidates: Dose, solubility and lipophilicity [J]. Drug Discovery Today, 2016, 21 (10): 1719-1727
- [84] Li J J, Zhang X J, Wang X H, et al. Discrimination of excess toxicity from baseline level for ionizable compounds: Effect of pH [J]. Chemosphere, 2016, 147: 382-388
- [85] Lin S Y, Yang X H, Liu H H. Development of liposome/ water partition coefficients predictive models for neutral and ionogenic organic chemicals [J]. Ecotoxicology and Environmental Safety, 2019, 179: 40-49
- [86] Escher B I, Schwarzenbach R P, Westall J C. Evaluation of liposome-water partitioning of organic acids and bases.
 2. Comparison of experimental determination methods [J]. Environmental Science & Technology, 2000, 34 (18): 3962-3968
- [87] Baumer A, Bittermann K, Klüver N, et al. Baseline toxicity and ion-trapping models to describe the pH-dependence of bacterial toxicity of pharmaceuticals [J]. Environmental Science Processes & Impacts, 2017, 19(7): 901-916
- [88] Schweigert N, Hunziker R W, Escher B I, et al. Acute toxicity of (chloro-) catechols and (chloro-) catechol-copper combinations in *Escherichia coli* corresponds to their membrane toxicity *in vitro* [J]. Environmental Toxicology and Chemistry, 2001, 20(2): 239-247
- [89] Klüver N, Bittermann K, Escher B I. QSAR for baseline toxicity and classification of specific modes of action of ionizable organic chemicals in the zebrafish embryo toxicity test [J]. Aquatic Toxicology, 2019, 207: 110-119
- [90] Escher B I, Baumer A, Bittermann K, et al. General baseline toxicity QSAR for nonpolar, polar and ionisable chemicals and their mixtures in the bioluminescence inhibition assay with *Aliivibrio fischeri* [J]. Environmental Science Processes & Impacts, 2017, 19(3): 414-428
- [91] Ng C A, Hungerbühler K. Bioaccumulation of perfluorinated alkyl acids: Observations and models [J]. Environ-

mental Science & Technology, 2014, 48(9): 4637-4648

- [92] Zhang K, Wiseman S, Giesy J P, et al. Bioconcentration of dissolved organic compounds from oil sands processaffected water by medaka (*Oryzias latipes*): Importance of partitioning to phospholipids [J]. Environmental Science & Technology, 2016, 50(12): 6574-6582
- [93] Timmer N, Droge S T J. Sorption of cationic surfactants to artificial cell membranes: Comparing phospholipid bilayers with monolayer coatings and molecular simulations [J]. Environmental Science & Technology, 2017, 51(5): 2890-2898
- [94] Endo S, Goss K U. Applications of polyparameter linear free energy relationships in environmental chemistry [J]. Environmental Science & Technology, 2014, 48 (21): 12477-12491
- [95] Abraham M H. Scales of solute hydrogen-bonding: Their construction and application to physicochemical and biochemical processes [J]. Chemical Society Reviews, 1993, 22(2): 73
- [96] Abraham M H, Ibrahim A, Zissimos A M. Determination of sets of solute descriptors from chromatographic measurements [J]. Journal of Chromatography A, 2004, 1037 (1-2): 29-47
- [97] Goss K U. Predicting the equilibrium partitioning of organic compounds using just one linear solvation energy relationship (LSER) [J]. Fluid Phase Equilibria, 2005, 233 (1): 19-22
- [98] Abraham M H, Zhao Y H. Determination of solvation descriptors for ionic species: Hydrogen bond acidity and basicity [J]. The Journal of Organic Chemistry, 2004, 69 (14): 4677-4685
- [99] Abraham M H, Zhao Y H. Characterisation of the water/ o-nitrophenyl octyl ether system in terms of the partition of nonelectrolytes and of ions [J]. Physical Chemistry Chemical Physics, 2005, 7(12): 2418-2422
- [100] Abraham M H, Acree W E. Equations for the transfer of neutral molecules and ionic species from water to organic phases [J]. The Journal of Organic Chemistry, 2010, 75 (4): 1006-1015
- [101] Abraham M H, Acree W E. The transfer of neutral molecules, ions and ionic species from water to wet octanol[J]. Physical Chemistry Chemical Physics, 2010, 12(40): 13182
- [102] Zhao Y F, Lin S, Choi J W, et al. Prediction of adsorption properties for ionic and neutral pharmaceuticals and pharmaceutical intermediates on activated charcoal from aqueous solution via LFER model [J]. Chemical Engineering Journal, 2019, 362: 199-206

- [103] Yu C L, Devlin J F, Bi E P. Bonding of monocarboxylic acids, monophenols and nonpolar compounds onto goethite [J]. Chemosphere, 2019, 214: 158-167
- [104] Enoch S J. The Use of Quantum Mechanics Derived Descriptors in Computational Toxicology [M]//Challenges and Advances in Computational Chemistry and Physics. Dordrecht: Springer Netherlands, 2009: 13-28
- [105] Zhang H B. A QSAR study of the brain/blood partition coefficients on the basis of pKa values [J]. QSAR & Combinatorial Science, 2006, 25(1): 15-24
- [106] Ding F, Yang X H, Chen G S, et al. Development of bovine serum albumin-water partition coefficients predictive models for ionogenic organic chemicals based on chemical form adjusted descriptors [J]. Ecotoxicology and Environmental Safety, 2017, 144: 131-137
- [107] Yang X H, Ou W, Xi Y, et al. Emerging polar phenolic disinfection byproducts are high-affinity human transthyretin disruptors: An *in vitro* and in silico study [J]. Environmental Science & Technology, 2019, 53 (12): 7019-7028
- [108] Xi Y, Yang X H, Zhang H Y, et al. Binding interactions of

halo-benzoic acids, halo-benzenesulfonic acids and halophenylboronic acids with human transthyretin [J]. Chemosphere, 2020, 242: 125135

- [109] Yang X H, Ou W, Zhao S S, et al. Rapid screening of human transthyretin disruptors through a tiered *in silico* approach [J]. ACS Sustainable Chemistry & Engineering, 2021, 9(16): 5661-5672
- [110] Yang X H, Ou W, Zhao S S, et al. Human transthyretin binding affinity of halogenated thiophenols and halogenated phenols: An *in vitro and in silico* study [J]. Chemosphere, 2021, 280: 130627
- [111] Mamy L, Patureau D, Barriuso E, et al. Prediction of the fate of organic compounds in the environment from their molecular properties: A review [J]. Critical Reviews in Environmental Science and Technology, 2015, 45 (12): 1277-1377
- [112] Bittermann K, Spycher S, Endo S, et al. Prediction of phospholipid-water partition coefficients of ionic organic chemicals using the mechanistic model COSMOmic [J]. The Journal of Physical Chemistry B, 2014, 118 (51): 14833-14842